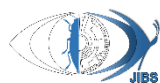


Original Article 

Macrofauna and mesofauna in Hyrcanian Protected Forest: Seasonal patterns of the arthropod diversity and community structure across soil depths

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ABSTRACT. Hyrcanian forests are considered among the oldest forests in the world, and due to their temperate and humid climate and fertile soil, they have rich fauna and flora. Soil-dwelling arthropods play essential roles in forest ecosystem functioning, contributing to litter decomposition, nutrient cycling, and soil structure maintenance. This study assessed the diversity and vertical distribution of macrofauna and mesofauna in Semeskandeh Forest, a protected Hyrcanian Forest in northern Iran, across different soil depths (leaf litter, 0–3 cm, 3–6 cm) and seasons over one year. Over 10,000 individuals were collected, representing diverse taxonomic groups, including ants, millipedes, springtails, mites, pseudoscorpions, and diplurans. Population densities and species diversity were consistently highest in the litter layer and declined with increasing soil depth, with clear seasonal peaks observed during the study period. Ants dominated the macrofauna, while mites and springtails were the most diverse mesofaunal taxa. The study revealed clear seasonal and vertical patterns in the density and diversity of soil faunal communities in the Hyrcanian forest ecosystem. These findings provide baseline information for biodiversity conservation and future monitoring of soil fauna in Hyrcanian forests. Thereby, conserving litter quality, maintaining habitat integrity, and monitoring soil faunal diversity are critical for sustaining forest ecosystem functions in such a globally valuable and inherited region.

KEYWORDS: Abundance, Community composition, Conservation, Faunal diversity, Iran, Leaf litter

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INTRODUCTION

Forests are among the most valuable natural resources on Earth and play an important role in climate regulation, provision of food and shelter, prevention of soil erosion, biodiversity conservation, regulation of surface water and groundwater recharge, soil protection and formation, as well as supporting industrial production (Ghomi-Avili et al. 2020; Zahriban Hesari et al. 2020; Truffer et al. 2025; Stritih et al. 2026). The Hyrcanian Forest, with its diverse vegetation, is recognized as one of the most valuable temperate broad-leaved deciduous forests in the world (Hosseini 2006; Haghdoust et al. 2011; Mehrafrooz Mayvan et al. 2015; D'Haese et al. 2025). It extends as a narrow green belt, 850 km in length and 20–70 km in width, from the southern shores of the Caspian Sea to northern Iran (Mohammadi et al. 2017; Waez-Mousavi 2017; Kooch et al. 2019; Homami Totmaj et al. 2021). These forests provide diverse habitats that support a wide range of soil invertebrate communities, linking the above-below ground ecosystem. Soil-dwelling arthropods are among the first organisms to have colonized terrestrial environments and are the most abundant animals on Earth, occurring in most

ecosystems. Their species richness is generally higher in forests than in other habitats (Zhang 2011; Eckert et al. 2023; Hébert 2023). Soil fauna plays crucial ecological roles in soil processes, food web dynamics, and overall ecosystem health (Chahartaghi et al. 2009; Domene 2016; Cayuela et al. 2020; Kooch & Noghre 2020; Ma et al. 2019; Coleman et al. 2023). Their contributions include litter decomposition, nutrient cycling (e.g., nitrogen and carbon), enhancement of soil quality and structure, serving as reliable indicators of soil environmental function, and indirectly contributing to the regulating microclimate and local hydrological processes, representing just a portion of the ecosystem services provided by arthropods (Guo et al. 2020; Lin et al. 2019; Berry et al. 2020; Domínguez et al. 2020; Gonçalves et al. 2020; Kooch et al. 2020; Menta & Remelli 2020; Sánchez-Montoya et al. 2020; Yang et al. 2020; Wang et al. 2021). Invertebrate communities that constitute the soil fauna are classified into three groups based on body size: macrofauna, mesofauna, and microfauna (Zagatto et al. 2017; Wu & Wang 2019; Marsden et al. 2020; Zagatto et al. 2020). Soil macrofauna are relatively large, with body sizes ranging from 2 to 20 mm (e.g., millipedes, insects, and earthworms). Mesofauna measure between 0.2 and 2 mm (e.g., springtails, mites, pseudoscorpions, some millipedes, oligochaetes, and several insect orders). Microfauna, including nematodes and protozoa, are smaller than 0.2 mm (Lavelle et al. 1997; Xin et al. 2012; García-Segura et al. 2018; Cayuela et al. 2020; Marsden et al. 2020; Zagatto et al. 2020).

Investigating soil-dwelling arthropod diversity provides valuable insights into environmental conditions because these organisms have limited mobility and highly specific microhabitat requirements (Heydari et al. 2020; Nacif et al. 2020). Sustainable forest management should prioritize biodiversity conservation. Human activities, including habitat loss and fragmentation combined with high pesticide inputs, are now recognized as major drivers of global biodiversity loss (Tilman et al. 2001; Tschamntke et al. 2005; de Pedro et al. 2020). The importance of biodiversity in supporting ecosystem services is increasingly acknowledged, and substantial experimental evidence demonstrates its positive influence on ecosystem functioning (Hector & Bagchi 2007; Tresch et al. 2019; O'Donnell & Wright 2021). Accurate characterization of biodiversity and a thorough understanding of dynamic ecological processes are both essential for the effective assessment and management of ecosystems (Decaëns et al. 2006; Valencia-Cuevas et al. 2020). In this context, understanding diversity patterns and their ecological relationships is particularly critical for conservation and for maintaining ecosystem functions (Xu et al. 2020).

The Hyrcanian forests in northern Iran exhibit complex patterns of species diversity and ecosystem structure that can be optimally managed only through comprehensive knowledge of this ecosystem. Although the biodiversity of vascular plants and vertebrates has been relatively well documented, recent studies indicate that soil arthropods, particularly Collembola, remain considerably understudied in these forests (Mehrafrouz Mayvan et al. 2015; D'Haese et al. 2025). Recent investigations have also demonstrated that Hyrcanian forests support highly diverse soil arthropod communities with a high proportion of endemic and potentially undescribed species, while substantial knowledge gaps remain regarding their diversity, community structure, spatial distribution, and ecological patterns across different habitats and protected areas (Shayanmehr et al. 2025; Tully et al. 2025). Consequently, below-ground taxa have received relatively limited ecological attention, highlighting the need for quantitative studies assessing soil faunal diversity across different soil depths and microhabitats to improve our understanding of forest ecosystem functioning and to support evidence-based conservation and sustainable forest management. Accordingly, the objectives of this study were to assess the seasonal variations in the density and diversity of soil-dwelling arthropods, investigate their vertical distribution across different soil depths, and evaluate the implications of these patterns for biodiversity conservation and sustainable ecosystem management in a protected Hyrcanian forest.

MATERIALS AND METHODS

Study Area. The study was conducted in the Semeskandeh Forest, a part of the Hyrcanian Forest in Mazandaran Province (Fig. 1). The Hyrcanian Forest forms a green belt along the northern slopes of the Alborz Mountains, extending to the southern coast of the Caspian Sea. It spans five provinces (North Khorasan, Golestan, Mazandaran, Gilan, and Ardabil), and covers approximately 1.9 million hectares up to 2800 m above sea level (Akhani et al. 2010; Nakhutsrishvili et al. 2015).

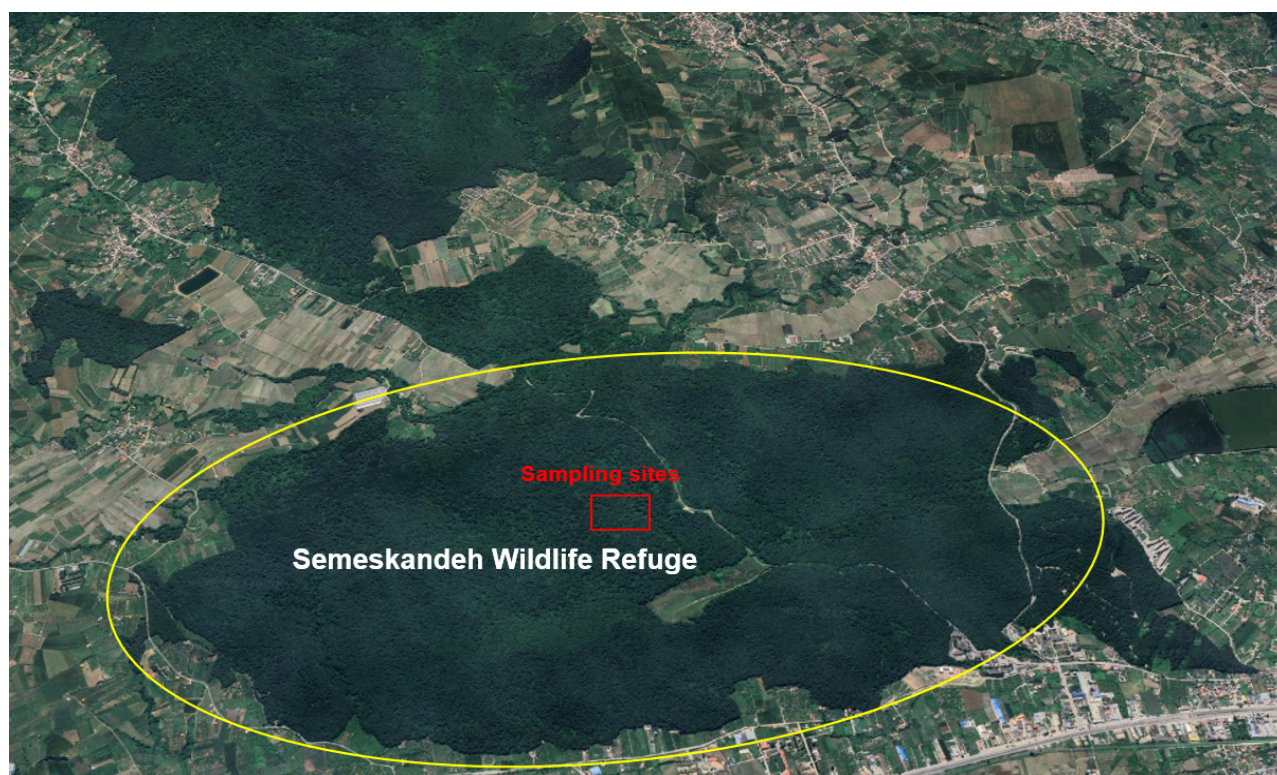


Figure 1. Semeskandeh forest in Mazandaran Province, northern Iran. The yellow outline indicates the boundary of the study forest, whereas the red rectangle indicates the sampling plots where soil arthropod samples were collected.

The forest exhibits diverse tree species, including Persian maple (*Acer velutinum* Boiss.), Cappadocian maple (*Acer cappadocicum* Gled.), Oriental beech (*Fagus orientalis* Lipsky), Caucasian alder (*Alnus subcordata* C.A.Mey.), European alder (*Alnus glutinosa* (L.) Gaertn.), chestnut-leaved oak (*Quercus castaneifolia* C.A.Mey.), Persian ironwood tree (*Parrotia persica* (DC.) C.A.Mey.), common hornbeam (*Carpinus betulus* L.), Scots elm (*Ulmus glabra* Huds.), field elm (*Ulmus minor* Mill.), white mulberry (*Morus alba* L.), European ash (*Fraxinus excelsior* L.), wild service tree (*Aria torminalis* (L.) Beck), Caucasian elm (*Zelkova carpinifolia* (Pall.) K.Koch), and Caucasian walnut (*Pterocarya fraxinifolia* (Lam.) Spach) (Gholizadeh et al. 2019; Ghorbanalizadeh & Akhani 2022). Semeskandeh Forest, covering approximately 1,041 hectares, is located in Mazandaran Province, 5 km southeast of Sari city along the Sari–Neka Road (36°32'–36°34'N, 53°07'–53°11'E) at an elevation of 50–90 m above sea level. The area has a temperate sub-humid climate, characterized by mild, humid winters and warm, relatively dry summers, with an average annual temperature of 15°C and mean annual precipitation of about 650 mm (Moradi et al. 2016). Soil types in the area include silty clay loam, silty loam, clay, silty clay, and clay loam, with a pH ranging from 6.2 to 7.9, close to neutral (Mehrafrooz Mayvan & Shayanmehr 2016).

Sampling methods. To sample soil-dwelling arthropods, specimens were collected from leaf litter (OL, F, H horizons) and soil at depths of 0–3 cm and 3–6 cm (Ah horizon). Eight samples, spaced at least 20 m apart, were taken in July, September, and November 2012, and in January, March, and May 2013, at two-month intervals. Samples for macrofauna were collected using a steel corer of 20 cm diameter (314 cm²), while mesofauna were sampled with a 5 cm diameter corer (19.625 cm²) (Kounda-Kiki et al. 2004; Chahartaghi et al. 2005, 2006; Palacios-Vargas et al. 2007). The corers were inserted into the soil, and the upper 10 cm of soil was collected and subsequently separated into the leaf litter, 0–3 cm, and 3–6 cm layers. Arthropods were extracted using a modified Berlese–Tullgren funnel consisting of individual plastic extraction units equipped with a wire mesh to support the soil samples and a 40-W incandescent lamp positioned above each sample to create a gradual heat and moisture gradient, forcing arthropods into collecting containers beneath the mesh, and were preserved in 75% ethanol. Specimens belonging to different groups were sent to the experts for precise species determination.

Taxonomic identification was carried out using available taxonomic keys, monographs, and regional faunal references relevant to each arthropod group. Insect larvae were treated as a separate category because they could not be reliably identified to order or family level. Specimens of Curculionidae (Coleoptera) were identified by Patrick Weil (Maison de la Nature et de l'Environnement, Pau, France), Carabidae (Coleoptera) by Ron Felix (Naturalis Biodiversity Center, Leiden, the Netherlands), Staphylinidae (Coleoptera) by Sinan Anlaş (Manisa Celal Bayar University, Manisa, Turkey), Formicidae (Hymenoptera) by Donat Agosti (Plazi, Bern, Switzerland), Isopoda by Stefano Taiti (Research Institute on Terrestrial Ecosystems, National Research Council, Florence, Italy), Diplopoda by Sergei Golovatch (Institute of Ecology and Evolution, Russian Academy of Sciences, Moscow, Russia), Diplura by Alberto Sendra (Universidad de Alcalá, Madrid, Spain), and Pseudoscorpiones by Mehdi Nasirkhani (Shahid Bahonar University of Kerman, Kerman, Iran). These taxa were identified to the species level. Collembola specimens were identified mainly using the keys provided by Bretfeld (1999), Potapov (2001), Thibaud et al. (2004), Fjellberg (2007), and Jordana (2012) and were identified to the species level. Due to the extremely high diversity and taxonomic complexity of mites (Acari), as well as limitations in available diagnostic resources for many groups, mites were identified mainly at the family level by Mahdi Jalaeian (Rice Research Institute of Iran, Rasht, Iran).

Statistical Analysis. The data were recorded in Microsoft Excel. Diversity of soil macrofauna and mesofauna was calculated according to Shannon–Wiener as $H' = -\sum p_i \ln p_i$, with p_i the ratio between the number of the i th species and the total number of species for each soil depth per sampling date (Pielou 1971; Ardakani 2007). Analysis of variance was used to analyze variations in the number of soil fauna species with sampling date and soil depth. Means were compared using the least significant difference (LSD) multiple range test at $p < 0.01$. Prior to statistical analysis, data were inspected for homogeneity of variance using the Levene test and log ($x+1$) transformed if necessary. Statistical analyses were performed using IBM SPSS Statistics version 21.0 (IBM Corp., Armonk, NY, USA) and Minitab Statistical Software version 16.0 (Minitab Inc., State College, PA, USA). The density values presented in Tables 3 and 4 represent monthly mean densities (mean $\pm$ SD) for each sampling month. The density ranges reported in the results refer to the minimum and maximum monthly mean densities observed during the sampling period.

RESULTS

Species composition

Soil macrofauna. A total of over 5,000 soil macrofauna individuals, representing various taxonomic groups, including insects (class Insecta), centipedes (class Chilopoda), millipedes (class Diplopoda), spiders (class Arachnida), and isopods (order Isopoda), were collected across different seasons and soil depths. Among them, Hexapods exhibited the highest density, followed by millipedes, while the lowest density was observed in spiders (Fig. 2). The class Insecta assemblage included several orders such as Blattodea, Heteroptera, Hymenoptera, and Coleoptera. Hymenoptera dominated the insect assemblage, with the family Formicidae accounting for 52% of individuals, followed by Coleoptera (44%), and Diptera (3%). Overall, a total of 70 soil-dwelling macrofauna species, belonging to 58 genera, 14 families, and 4 orders, were collected and identified from Semeskandeh Forest (Table 1).

Soil mesofauna. A total of over 6,400 mesofauna individuals were collected. The mesofauna community included mites (subclass Acari), springtails (class Collembola), proturans (class Protura), diplurans (class Diplura), psocids (class Insecta: order Psocoptera), and pseudoscorpions (class Arachnida: order Pseudoscorpiones). Among these groups, mites (45%) and springtails (40%) were the most abundant taxa (Fig. 3). Within subphylum Hexapoda, a total of 19 Collembola species, belonging to 17 genera and 10 families, were identified (Table 2). Additionally, 2 Diplura species (1 genus, 2 subgenera, 2 species) and 3 Psocoptera species (3 genera, 3 species) were recorded. Excluding Hexapoda, 4 pseudoscorpion species were collected. Also, mites comprised 33 species from 16 families within the orders Oribatida (superorder Acariformes), Mesostigmata (superorder Parasitiformes) and Trombidiformes (superorder Acariformes). The families within the Mesostigmata order included Uropodidae, Macrochelidae, Rhodacaridae, Phytoseiidae, Veigaiidae, Pachylaelapidae, Laelapidae, Parholaspididae, and Epicroiidae.

The families within the Trombidiformes order included Trombidiidae, Cunaxidae, Pseudocheylidae, Bdellidae, Erythraeidae, Labidostomatidae, and Caligonellidae. Within Mesostigmata, Uropodidae was the most abundant family (16%), while Epicriidae was the least abundant (1%) (Fig. 4). In Trombidiformes, Cunaxidae showed the highest abundance (17%), whereas Labidostomatidae was the least numerous (1%).

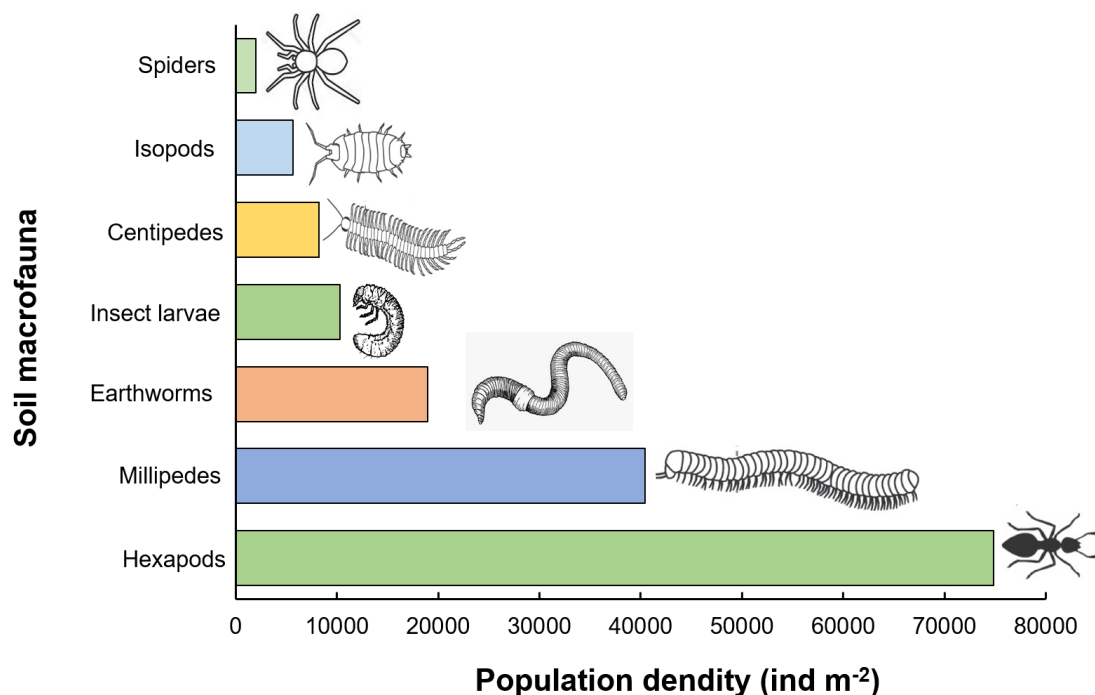


Figure 2. Composition and population density of macrofauna collected from different soil layers in Semeskandeh Forest (Mazandaran Province, Iran).

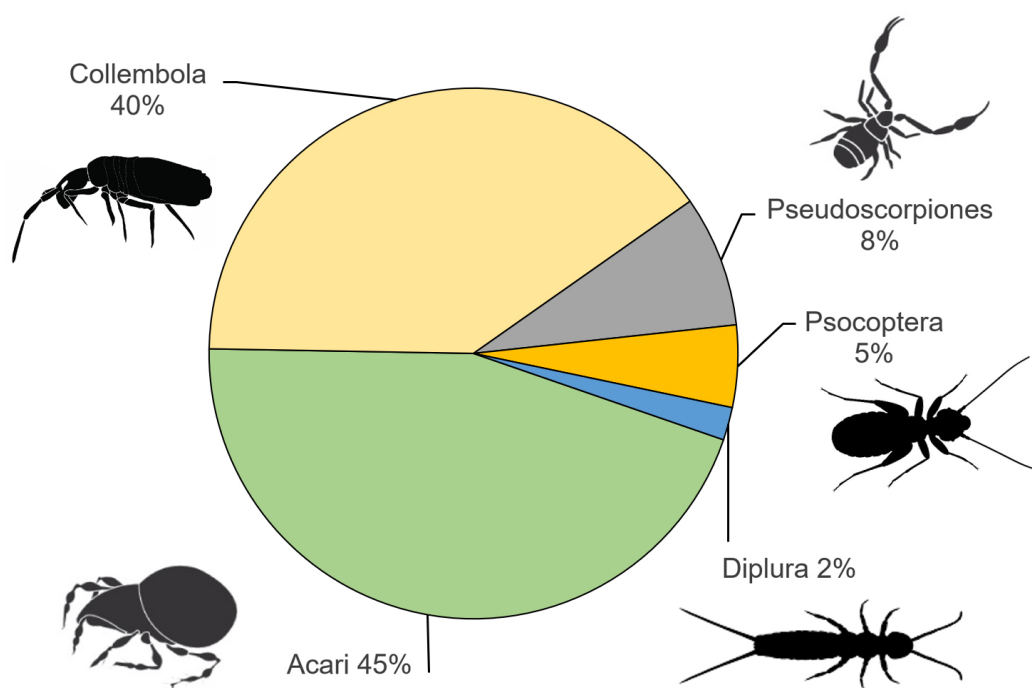


Figure 3. Abundance percentage of soil mesofauna collected from different layers of Semeskandeh Forest soil.

Table 1. Soil macrofauna recorded in the Semeskandeh Forest (Mazandaran Province, Iran).

Categories	Family	Species
Class Chilopoda	Dignathodontidae	<i>Henia bicarinata</i> (Meinert, 1870)
	Lithobiidae	<i>Lithobius chalusensis</i> Matic, 1969
		<i>Lithobius forficatus</i> (Linnaeus, 1758)
Class Diplopoda	Blaniulidae	<i>Nopoiulus iranica</i> Enghoff, 1984
	Julidae	<i>Amblyiulus</i> sp.
		<i>Cylindroiulus treptoflagellum</i> Read, 1992
		<i>Leptoiulus</i> sp.
	Lophoproctidae	<i>Lophoproctus lucidus</i> (Chalande, 1888)
	Paradoxosomatidae	<i>Strongylosoma kordylamythrum</i> Attems, 1898
	Polydesmidae	<i>Brachydesmus ferrugineus</i> Lohmander, 1936
Order Isopoda	Porcellionidae	<i>Brevurus masandaranus</i> Schmalfuss, 1986
	Trachelipodidae	<i>Trachelipus pieperi</i> Schmalfuss, 1986
Order Coleoptera	Attelabidae	<i>Neocoenorrhinus</i> (<i>Schoenitemnus</i>) <i>minutus</i> (Herbst, 1797)
	Carabidae	<i>Acinopus laevigatus</i> (Ménétriés, 1832)
		<i>Acupalpus maculatus</i> (Schaum, 1860)
		<i>Amara aenea</i> (De Geer, 1774)
		<i>Amara apricaria</i> (Paykull, 1790)
		<i>Amara familiaris</i> (Duftschmidt, 1812)
		<i>Amara ovata</i> (Fabricius, 1792)
		<i>Anchomenus dorsalis</i> (Pontoppidan, 1763)
		<i>Bembidion</i> (<i>Phyla</i>) <i>incommodum</i> Netolitzky, 1926
		<i>Brachinus brevicollis</i> (Motschulsky, 1844)
		<i>Broscus karelini</i> (Zoubkoff, 1837)
		<i>Carabus</i> (<i>Shodristocarabus</i>) <i>adamsi</i> (Adams, 1817)
		<i>Chlaenius festivus</i> (Panzer, 1796)
		<i>Chlaenius tristis</i> (Schaller, 1783)
		<i>Cicindela monticola</i> Ménétriés, 1832
		<i>Cylindera germanica</i> (Linnaeus, 1758)
		<i>Cymindis andreae</i> (Ménétriés, 1832)
		<i>Distichus planus</i> (Bonelli, 1813)
		<i>Ditomis calydonius oriens</i> (Dvorak, 1993)
		<i>Harpalus convexus</i> Faldermann, 1836
		<i>Harpalus rufipes</i> (De Geer, 1774)
		<i>Harpalus smyrnensis medicus</i> (Kataev, 1993)
		<i>Odontocarus asiaticus</i> (Chaudoir, 1852)
		<i>Parophonus planicollis</i> (Dejean, 1829)
		<i>Poecilus cupreus</i> (Linnaeus, 1758)
		<i>Sphaerotachys hoemorrhoidalis</i> (Ponza, 1805)
		<i>Stenolophus proximus</i> (Dejean, 1829)
		<i>Stenolophus abdominalis persicus</i> Mannerheim, 1844
		<i>Syntomus obscuriguttatus</i> (Duftschmidt, 1812)
		<i>Trechus quadristriatus</i> (Schrank, 1781)
	Curculionidae	<i>Asproparthenis punctiventris</i> (Germar, 1824)
		<i>Bagous bagdatensis</i> Pic, 1904
		<i>Larinus centaurea</i> Oliver, 1807
		<i>Lixus angustatus</i> (Fabricius, 1775)
		<i>Lixus</i> (<i>Broconius</i>) <i>rubicundus</i> Zoubkoff, 1833
		<i>Polydrusus</i> (<i>Scythodrusus</i>) <i>inustus</i> Germar, 1824
		<i>Sitophilus granarius</i> (Linnaeus, 1758)
		<i>Sitophilus oryzae</i> Hustache, 1930
		<i>Sitophilus zeamais</i> (Motschulsky, 1855)
		<i>Sternuchopsis karelini</i> (Boheman, 1844)
	Staphylinidae	<i>Physetops tataricus</i> (Pallas, 1773)
		<i>Gauropterus sanguinipes</i> (Reitter, 1889)
		<i>Ocypus curtipennis</i> Motschulsky, 1849
		<i>Quedius mixtus</i> Eppelsheim, 1890
		<i>Paederus balcanicus</i> Koch, 1938
		<i>Astenus klapperichi</i> Coiffait, 1980

Categories	Family	Species
Order Hymenoptera	Formicidae	<i>Aphaenogaster subterranea</i> (Latreille, 1798)
		<i>Formica cunicularia</i> Latreille, 1798
		<i>Formica sanguinea</i> Latreille, 1798
		<i>Lepisiota semenovi</i> (Ruzsky, 1905)
		<i>Messor denticulatus</i> Santschi, 1927
		<i>Myrmica bergi</i> Ruzsky, 1902
		<i>Paratrechina longicornis</i> (Latreille, 1802)
		<i>Pheidole pallidula</i> (Nylander, 1849)
		<i>Ponera</i> sp.
		<i>Tapinoma karavaievi</i> Emery, 1925
		<i>Temnothorax parvulus</i> (Schenck, 1852)

Table 2. Soil mesofauna recorded in the Semeskandeh Forest (Mazandaran Province, Iran).

Category	Family	Species
Order Pseudoscorpiones	Chthoniidae	<i>Chthonius (Chthonius) shelkovnikovi</i> Redikorzev, 1930
		<i>Chthonius (Ephipiochthonius) tetrachelatus</i> (Preyssl, 1790)
	Neobisiidae	<i>Roncus corimanus</i> Beier, 1951
		<i>Roncus viti</i> Mahnert, 1974
Class Collembola	Dicyrtomidae	<i>Dicyrtoma ghilarovi</i> Bretfeld, 1996
		<i>Dicyrtoma</i> sp.
	Entomobryidae	<i>Heteromurus major</i> (Moniez, 1889)
		<i>Heteromurus nitidus</i> (Templeton, 1836)
		<i>Lepidocyrtus</i> sp.
		<i>Pseudosinella octopunctata</i> Börner, 1901
	Hypogastruridae	<i>Ceratophysella stercoraria</i> Stach, 1963
	Isotomidae	<i>Folsomia penicula</i> Bagnall, 1939
		<i>Isotomiella minor</i> (Schäffer, 1896)
		<i>Isotomurus italicus</i> Carapelli, Frati, Fanciulli & Dallai, 1995
	Katiannidae	<i>Sminthurinus elegans</i> (Fitch, 1863)
	Neanuridae	<i>Persanura hyrcanica</i> Mehrafrooz Mayvan, Shayanmehr, Smolis & Skarżyński, 2015
	Onychiuridae	<i>Heteraphorura japonica</i> (Yosii, 1967)
	Sminthuridae	<i>Caprainea marginata</i> (Schött, 1893)
		<i>Lipothrix lubbocki</i> (Tullberg, 1872)
		<i>Paralipothrix cf. natalicus</i> (Ellis, 1974)
		<i>Sminthurus ghilarovi</i> Stebaeva, 1966
		<i>Sphaeridia</i> sp.
	Tomoceridae	<i>Tomocerus vulgaris</i> (Tullberg, 1871)
Class Diplura	Campodeidae	<i>Campodea (Campodea) fragilis</i> Meinert, 1865
		<i>Campodea (Dicampa) sprovieri</i> Silvestri, 1932
Order Psocoptera	Ectopsocidae	<i>Ectopsocopsis cryptomeriae</i> (Enderlein, 1907)
	Trogiidae	<i>Lepinotus reticulatus</i> Enderlein, 1904
	Liposcelididae	<i>Liposcelis paeta</i> Pearman, 1942

Population density and biodiversity

Macrofauna. The total density of soil macrofauna ranged from 32,746 to 42,312 individuals m⁻² in the litter layer, from 11,776 to 14,408 individuals m⁻² in the 0–3 cm soil depth, and from 7,784 to 12,267 individuals m⁻² in the 3–6 cm soil depth (Table 3). In the litter layer, the highest density was observed in autumn, particularly in September. After January, macrofaunal density increased with rising temperatures, reaching a peak in September. In the 0–3 cm layer, the highest densities were recorded in July and November, with numbers remaining relatively stable throughout the year, except for minor increases in July and November. The 3–6 cm layer, which had the lowest densities among the three layers, reached its highest density in January. The highest diversity value was recorded in the litter layer during autumn (September) (Table 3), whereas the lowest value occurred in the 3–6 cm soil depth in March. Throughout the year, diversity values in the litter layer were consistently higher than in the two deeper layers, except in January, when the 0–3 cm layer exhibited a slightly higher diversity value than the litter layer.

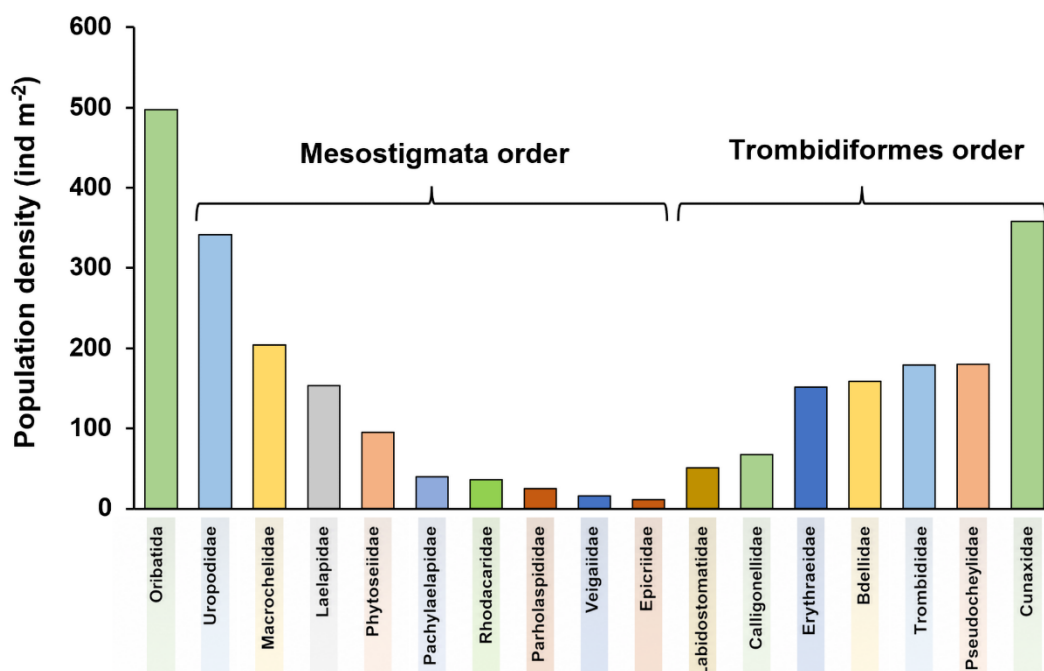


Figure 4. Population density of soil mites (Oribatida, Mesostigmata and Trombidiformes orders) in Semeskandeh Forest.

Table 3. Seasonal changes in density and diversity of soil macrofauna at three soil depths recorded in the Semeskandeh Forest (means $\pm$ SD).

	January	March	May	July	September	November	Average
Density (ind. m⁻²)							
Leaf litter	1831 $\pm$ 455	820 $\pm$ 67	764 $\pm$ 46	2512 $\pm$ 534	3213 $\pm$ 1335	1429 $\pm$ 360	1761 $\pm$ 466
0-3 cm	860 $\pm$ 215	788 $\pm$ 199	776 $\pm$ 110	1441 $\pm$ 458	951 $\pm$ 188	1441 $\pm$ 458	1042 $\pm$ 271
3-6 cm	784 $\pm$ 250	243 $\pm$ 51	271 $\pm$ 40	426 $\pm$ 183	701 $\pm$ 279	744 $\pm$ 209	528 $\pm$ 168
Total	3475 $\pm$ 920	1851 $\pm$ 317	1811 $\pm$ 196	4379 $\pm$ 1175	4565 $\pm$ 1802	3614 $\pm$ 1027	3331 $\pm$ 905
Diversity (<i>H'</i>)							
Leaf litter	1.48 $\pm$ 0.13	1.91 $\pm$ 0.33	1.99 $\pm$ 0.31	1.85 $\pm$ 0.31	2.32 $\pm$ 0.23	1.79 $\pm$ 0.25	1.89 $\pm$ 0.29
0-3 cm	1.54 $\pm$ 0.14	1.32 $\pm$ 0.41	1.51 $\pm$ 0.33	1.83 $\pm$ 0.37	1.79 $\pm$ 0.41	1.60 $\pm$ 0.27	1.60 $\pm$ 0.34
3-6 cm	1.10 $\pm$ 0.17	0.95 $\pm$ 0.27	1.41 $\pm$ 0.31	1.22 $\pm$ 0.59	1.83 $\pm$ 0.51	1.13 $\pm$ 0.54	1.27 $\pm$ 0.43
Total	4.13 $\pm$ 0.44	4.19 $\pm$ 1.01	4.92 $\pm$ 0.95	4.91 $\pm$ 1.27	5.96 $\pm$ 1.15	4.53 $\pm$ 1.06	4.77 $\pm$ 1.07

Table 4. Seasonal changes in density and diversity of soil mesofauna at three soil depths recorded in the Semeskandeh Forest (means $\pm$ SD).

	January	March	May	July	September	November	Average
Density (ind. m⁻²)							
Leaf litter	847 $\pm$ 116	1070 $\pm$ 160	903 $\pm$ 99	995 $\pm$ 177	1078 $\pm$ 164	839 $\pm$ 211	988 $\pm$ 154
0-3 cm	473 $\pm$ 125	672 $\pm$ 80	637 $\pm$ 123	453 $\pm$ 107	720 $\pm$ 176	417 $\pm$ 93	562 $\pm$ 117
3-6 cm	441 $\pm$ 90	477 $\pm$ 88	445 $\pm$ 88	238 $\pm$ 88	350 $\pm$ 67	417 $\pm$ 93	395 $\pm$ 86
Total	1763 $\pm$ 332	2221 $\pm$ 329	1986 $\pm$ 311	1687 $\pm$ 372	2149 $\pm$ 408	1675 $\pm$ 398	1913 $\pm$ 358
Diversity (<i>H'</i>)							
Leaf litter	2.29 $\pm$ 0.13	2.38 $\pm$ 0.21	2.29 $\pm$ 0.13	2.11 $\pm$ 0.38	2.34 $\pm$ 0.11	2.24 $\pm$ 0.21	2.27 $\pm$ 0.19
0-3 cm	2.03 $\pm$ 0.14	2.21 $\pm$ 0.17	2.02 $\pm$ 0.43	1.88 $\pm$ 0.33	2.29 $\pm$ 0.31	1.97 $\pm$ 0.19	2.06 $\pm$ 0.26
3-6 cm	2.01 $\pm$ 0.17	2.14 $\pm$ 0.18	2.11 $\pm$ 0.23	1.59 $\pm$ 0.35	1.51 $\pm$ 0.15	1.97 $\pm$ 0.19	1.88 $\pm$ 0.21
Total	6.33 $\pm$ 0.44	6.73 $\pm$ 0.56	6.42 $\pm$ 0.79	5.5 $\pm$ 1.06	6.14 $\pm$ 1.57	6.18 $\pm$ 0.59	6.23 $\pm$ 0.66

Among the soil macrofauna, ants (family Formicidae) exhibited the highest population density and community diversity across all soil depths. Their density ranged approximately from 120 to 1,470 individuals m^{-2} in the litter layer, from 175 to 400 individuals m^{-2} in the 0–3 cm soil depth, and from 110 to 690 individuals m^{-2} in the 3–6 cm layer. The highest densities were recorded in the litter layer during autumn, particularly in September (Fig. 5A), after which it declined with decreasing temperatures. The lowest densities across the three soil depths occurred in the 3–6 cm layer during spring (May).

Diversity showed seasonal variation, with the highest values observed in the litter layer in September and the lowest in the 0–3 cm layer during winter (January). Throughout the year, diversity in the litter layer remained higher than in the deeper soil depths, except in January (winter), when the 0–3 cm layer exhibited a slightly higher value than the litter layer (Fig. 5B). Millipedes exhibited variable densities across soil depths. Their densities ranged approximately from 220 to 770 individuals m^{-2} in the litter layer, from 170 to 740 individuals m^{-2} in the 0–3 cm soil depth, and from 90 to 150 individuals m^{-2} in the 3–6 cm depth (Fig. 5C). Peaks of density were recorded in July during the summer, while density decreased significantly in May (Fig. 5C). The results indicate that millipede species in this study remained active throughout the year in Semeskandeh Forest. Vertical distribution analysis showed that millipedes were relatively evenly distributed across all soil depths during different seasons, although their density was generally higher in the litter layer, except in November, when densities in the 0–3 cm layer exceeded those in the litter and 3–6 cm layers. Diversity followed a similar pattern, with the highest values observed in the litter layer in September and the lowest in the 3–6 cm layer in January (Fig. 5D). Soil-dwelling centipedes exhibited densities ranging from approximately 51.41 to 142.08 individuals m^{-2} in the litter layer, from 56.34 to 120.03 individuals m^{-2} in the 0–3 cm soil depth, and from 29.39 to 51.44 individuals m^{-2} in the 3–6 cm layer. Peak densities occurred in September during autumn, whereas density declined significantly in March and May. The results also indicated that the three species recorded in this study were active throughout the year in Semeskandeh Forest. Vertical distribution analysis showed that centipedes were relatively evenly distributed across all soil depths throughout the seasons, although they generally showed higher densities in the litter layer, except in July and January, when densities in the 0–3 cm layer exceeded those in the litter and 3–6 cm layers (Fig. 5E). Diversity patterns mirrored these trends, with the highest value observed in the 0–3 cm layer in July (summer) and the lowest in the 3–6 cm layer during the same month. Overall, diversity in the litter layer remained higher than in the other layers throughout most of the year, except in January and July, when the 0–3 cm layer exhibited slightly higher diversity than the litter layer (Fig. 5F).

Mesofauna. Densities of soil mesofaunal arthropods ranged from 839.81 to 1,077.66 individuals m^{-2} in the litter layer, from 417.83 to 720.38 individuals m^{-2} in the 0–3 cm soil depth, and from 238.85 to 477.71 individuals m^{-2} in the 3–6 cm layer (Table 4). Overall, mesofaunal arthropod densities remained relatively stable across seasons in Semeskandeh Forest. Peak densities were recorded in the litter layer and in the 0–3 cm soil depth during autumn (September). In the 0–3 cm soil depth, densities showed moderate seasonal fluctuations, with the highest value recorded in September and the lowest in November. The lowest densities among the three soil depths were observed in the 3–6 cm depth during July (summer). The diversity of mesofaunal arthropods ranged from 2.11 to 2.38 individuals m^{-2} in the litter layer, from 1.88 to 2.29 individuals m^{-2} in the 0–3 cm soil depth, and from 1.51 to 2.14 individuals m^{-2} in the 3–6 cm soil depth (Table 4). The highest diversity value was observed in the litter layer during March and September, whereas the lowest diversity occurred in the 0–3 cm soil depth in July. Throughout the year, diversity values in the litter layer were consistently higher than those in the deeper soil depths and remained relatively stable across all months. The density and diversity of the most abundant mesofaunal groups, including Collembola, Acari, Diplura, and Pseudoscorpiones, were further examined across different soil depths and seasons.

Density of Collembola ranged approximately from 1,800 to 5,500 individuals m^{-2} in the litter layer, 1,250 to 2,500 individuals m^{-2} in the 0–3 cm depth, and 750 to 3,100 individuals m^{-2} in the 3–6 cm depth. Their density peaked during the cold seasons, with the highest density recorded in the litter layer in winter (January). In the 0–3 cm depth, the maximum density was also observed in January, while overall fluctuations remained relatively stable throughout the year, except for a decline in May, followed by an increase in July and September (Fig. 6A).

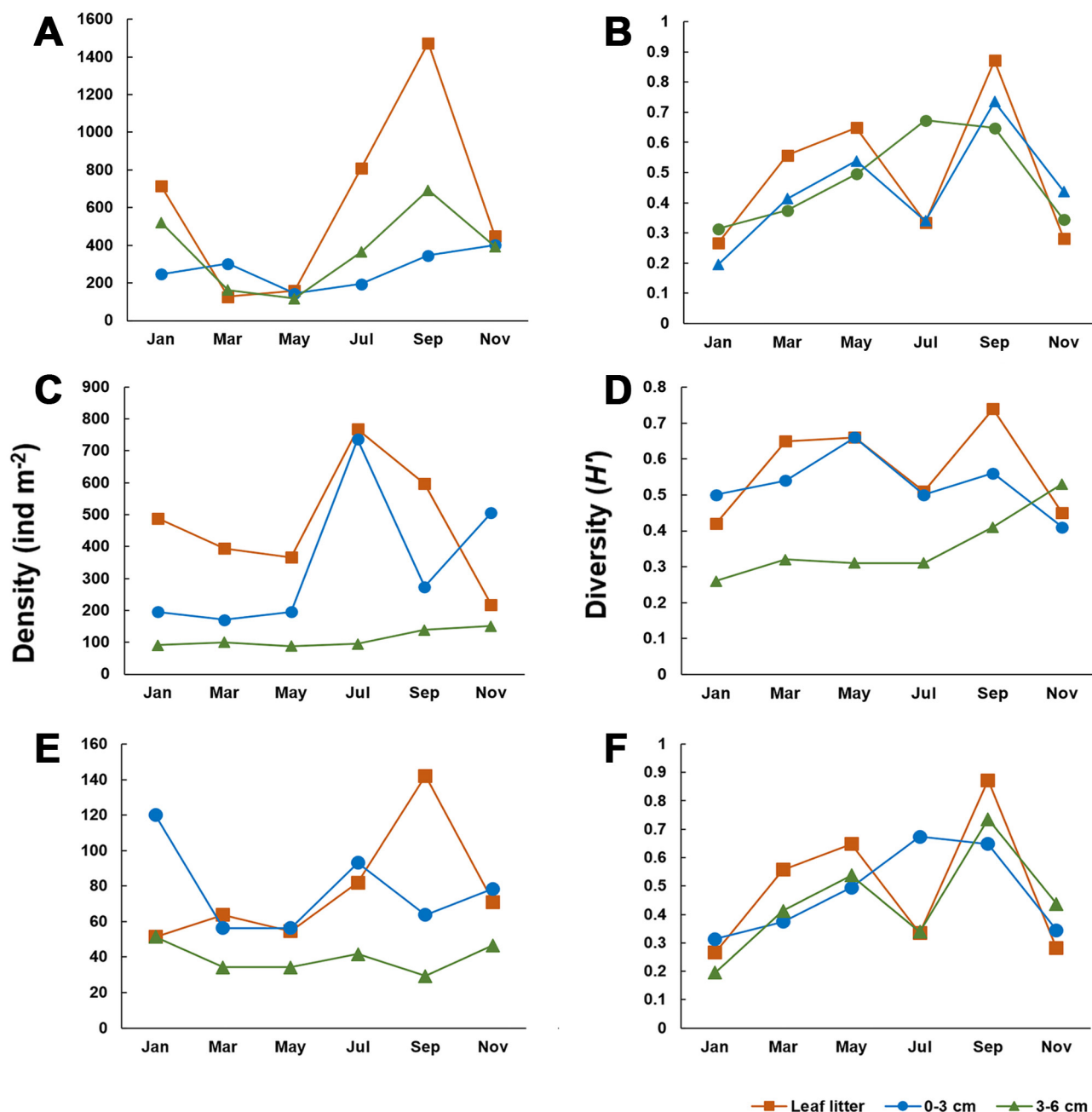


Figure 5. Seasonal changes in density (ind. m⁻²) (A, C, E), and diversity (B, D, F) at different soil depths. **A–B.** Formicidae; **C–D.** Diplopoda; **E–F.** Chilopoda. Samples were taken at bimonthly intervals in the leaf litter (OL, F, H), 0–3 and 3–6 cm soil depths.

The lowest density across all depths occurred in the 3–6 cm depth during July (summer). In general, higher densities in the 0–3 cm depth were associated with cold and humid periods, particularly in January and September. Also, community diversity of springtails ranged from 0.86 to 1.50 in the litter layer, 0.61 to 1.07 in the 0–3 cm layer, and 0.55 to 1.04 in the 3–6 cm depth. The highest diversity value was recorded in the litter layer during January and March, whereas the lowest value occurred in the 0–3 cm depth in July (Fig. 6B). Overall, diversity in the litter layer remained relatively stable throughout the year, except in September, when the 0–3 cm depth exhibited a slightly higher diversity value. Density of soil-dwelling mites ranged approximately from 770 to 2056 individuals m⁻² in the litter layer, 573 to 1046 individuals m⁻² in the 0–3 cm depth, and 425 to 861 individuals m⁻² in the 3–6 cm depth. Density remained relatively stable across months, with the highest values recorded in the litter layer during March and September (Fig. 6C). In the 0–3 cm depth, density peaked in September and was lowest in May, while in the 3–6 cm depth the lowest density occurred in January.

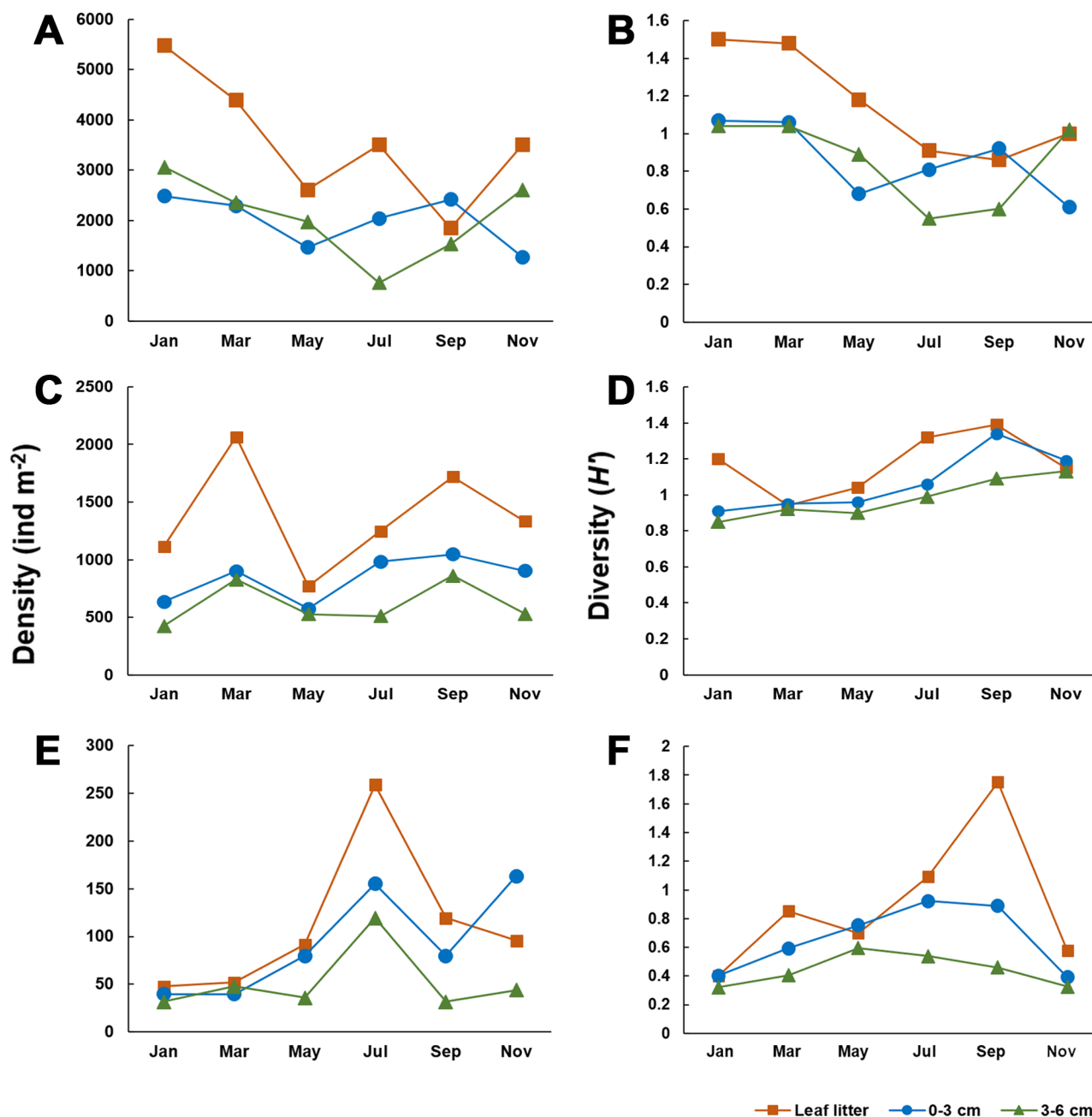


Figure 6. Seasonal changes in density (ind. m⁻²) (A, C, E), and diversity (B, D, F) at different soil depths. **A–B.** Collembola; **C–D.** Acari; **E–F.** Pseudoscorpiones. Samples were taken at bimonthly intervals in the leaf litter (OL, F, H), 0–3 and 3–6 cm soil depths.

The diversity value of soil mites was highest in September in the litter and 0–3 cm depths, whereas the lowest value was recorded in the 3–6 cm depth in January. In this deeper depth, diversity remained almost constant throughout the year (Fig. 6D). The highest density of pseudoscorpions was recorded in the litter layer in July (Fig. 6E). In the litter layer, their density peaked in July, while the lowest densities were observed in January. Following May, density increased sharply, reaching a peak in July. After September, densities declined toward November. A similar seasonal trend was observed in both the 0–3 cm and 3–6 cm soil depths, with the lowest values in January and March. The highest diversity of pseudoscorpion assemblages was observed in the litter layer during September, while the lowest was recorded in the 3–6 cm depth in January (Fig. 6F). Community diversity increased gradually from March, reaching its peak in September in the litter layer, and then declined again with decreasing temperatures. In contrast, the 0–3 cm and 3–6 cm depths showed no significant differences in diversity across months, maintaining a relatively stable trend throughout the year (Fig. 6F).

DISCUSSION

Over 10,000 individuals of various soil-dwelling animal groups were collected in the present study. Our results align with previous studies, showing that the highest densities occurred in the litter layer. The litter plays a pivotal role for soil invertebrates' populations, functioning as a protective cover against erosion, regulating soil moisture and temperature, and serving as a major source of organic matter (Gomyo & Kuraji 2016; Li et al. 2017; Marsden et al. 2020; Zhang et al. 2023; Liang 2024). In general, old-growth forests, due to their structural complexity, richer litter composition, and favorable microclimatic conditions, provide habitats that support higher soil faunal density and diversity (Burton & Eggleton 2016; Pereira et al. 2017; Pestana et al. 2020; Venuste 2020; Adiningrat et al. 2025). In particular, mites and collembolans were the most diverse mesofaunal taxa, and ants were dominant in the macrofauna community (Takeda & Abe 2001; Gagnarli et al. 2015; Ghiglieno 2020). The high density of ants may be attributed to their notable tolerance to environmental stress and even chemical pollution, allowing them to maintain large populations under varying soil conditions (Hasin & Booncher 2020). Semeskandeh Forest, as part of the ancient Hyrcanian forests, has been protected for more than four decades with limited human disturbance. Comparable studies on seasonal and vertical distribution patterns of both soil macrofauna and mesofauna in the Hyrcanian forests remain very limited. Therefore, the present study provides baseline information on the composition, density, diversity, and seasonal dynamics of soil arthropods in this unique forest ecosystem. Nevertheless, anthropogenic impacts such as industrial logging, heavy machinery traffic, agricultural expansion, and chemical inputs remain important pressures in the Hyrcanian forests and may adversely affect soil fauna communities (Jourgholami et al. 2019).

The present study revealed clear seasonal and vertical patterns in the density and diversity of soil fauna across different soil depths. Previous studies have shown that environmental factors such as vegetation type, rainfall, soil pH, and microclimate can influence the density and composition of soil fauna (Kelleher & Choi 2020). Plant cover has also been reported to affect both resource availability and habitat structure, with changes in plant community composition leading to shifts in soil arthropod assemblages (Kelleher & Choi 2020). Likewise, diverse vegetation generally supports more abundant and diverse soil fauna by providing food resources and suitable habitats for reproduction (Utami et al. 2019). Seasonal precipitation and soil moisture are among the major drivers of soil faunal dynamics, particularly for mesofaunal groups such as collembolans, which often reach peak densities during wetter periods, whereas mite populations may decline under dry conditions (Wiwatwitaya & Takeda 2005; Palacios-Vargas et al. 2007). Extreme temperature fluctuations may also influence soil fauna, especially epigeic taxa inhabiting the litter layer. In addition, population dynamics are often density-dependent, with periods of high density followed by declines associated with food limitation, intra-specific competition, predation, or parasitism (Sibly & Hone 2002; Pequeno et al. 2020). According to previous studies, soils in Semeskandeh Forest are slightly acidic to nearly neutral (pH 6.2–7.9), a condition that may influence the distribution of certain taxa, as some groups, such as isopods, are known to prefer neutral to alkaline soils (van Straalen 1998).

The seasonal density patterns observed in the present study suggest that vertical redistribution may have occurred in some soil arthropod groups in response to seasonal changes in environmental conditions. In particular, *Collembola* exhibited the clearest pattern, with a marked decline in litter layer density during September accompanied by a relative increase in the 0–3 cm soil depth, which may reflect a downward movement in response to changes in soil temperature or moisture. A similar but less consistent pattern was observed for *Chilopoda*. In contrast, *Acari*, *Pseudoscorpiones*, *Diplopoda*, and *Formicidae* generally showed parallel seasonal fluctuations across soil depths rather than reciprocal changes between adjacent horizons, providing little evidence for consistent vertical migration. Overall, these observations suggest that vertical redistribution, where present, was taxon-specific and likely represented a behavioral response of certain groups rather than a general pattern across the entire soil fauna community. Beyond these seasonal shifts, vertical stratification further shaped community patterns, with both density and diversity of soil fauna decreasing sharply with increasing soil depth. The litter layer, as the main entry point of organic matter into the soil, sustains the highest densities. While the reduction in organic matter content with depth partly explains this pattern, other factors

such as microclimatic shifts and habitat constraints also contribute (Berg 2012; Moradi et al. 2020). Interestingly, in line with Frouz et al. (2004), our results showed that the diversity of some soil faunal groups increased in deeper depths during colder months, supporting the taxon-specific patterns of vertical redistribution observed in the present study rather than a general refuge strategy across the entire soil fauna community.

Beyond biodiversity assessment, the observed seasonal and vertical patterns of soil fauna are important for forest soil health and sustainable soil management. Soil fauna contribute to litter fragmentation, decomposition, nutrient cycling, and microbial activity, which helps maintain soil structure and soil ecological functions (Zan et al. 2022; Zhang et al. 2023; Angst et al. 2024). In forest ecosystems, these processes are closely linked to carbon cycling because soil fauna can affect both the release of carbon and nitrogen from decomposing litter and the formation and stability of more readily decomposed and more stable soil organic matter (Anthony et al. 2023; Zhang et al. 2023; Angst et al. 2024). Therefore, conserving litter layers and reducing human disturbances such as soil compaction, intensive logging, and litter removal should be considered important management strategies for maintaining soil faunal diversity, soil health, and the carbon storage potential of Hyrcanian forest soils. Future studies should integrate soil faunal diversity with measurements of soil organic carbon fractions, microbial community structure, enzyme activities, soil respiration, and physicochemical soil properties to better clarify the mechanisms linking soil fauna, soil health, and carbon storage under seasonal and environmental changes.

Considering the key role of soil fauna in maintaining ecosystem health, assessing changes in soil faunal diversity and species composition in the valuable Hyrcanian forests is of great importance. In the present study, soil samples were collected across different soil depths and seasons over one year, revealing that even small volumes of soil harbor numerous macrofaunal and mesofaunal groups at various depths. The protected status of Semeskandeh Forest and the lack of human disturbance significantly contributed to high density and diversity values. In contrast, most Hyrcanian forests remain unprotected and are heavily impacted by human activities, which can severely reduce soil fauna populations and disrupt ecosystem balance. The observed seasonal patterns suggest that changes in soil moisture and temperature may have influenced the population dynamics of forest soil fauna. Previous studies have shown that forest age, management practices, and vegetation diversity directly influence soil fauna density and diversity, which in turn play a crucial role in improving soil physicochemical properties, promoting plant growth, and sustaining overall forest ecosystem health. Given the ecological importance of soil fauna and the vulnerability of unprotected forests, targeted efforts to study and conserve these organisms and their habitats are essential.

AUTHOR CONTRIBUTIONS

The authors confirm their contribution to the paper as follows: M. Mehrafrooz Mayvan: Investigation, Methodology, Data curation, Formal analysis, Visualization, and Writing – original draft; M. Shayanmehr: Conceptualization, Methodology, Supervision, Validation, Project administration, and Writing – review & editing. The authors read and approved the final version of the manuscript.

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AVAILABILITY OF DATA AND MATERIAL

All source data used in this study are cited in the Materials and Methods section. The specimens listed in this study are deposited in Sari University of Agricultural Sciences and Natural Resources (SANRU) and are available from the curator upon request.

ETHICS APPROVAL AND CONSENT TO PARTICIPATE

This study only included arthropod material, and all required ethical guidelines for the treatment and use of animals were strictly adhered to in accordance with international, national, and institutional regulations. No human participants were involved in any studies conducted by the authors for this article.

CONSENT FOR PUBLICATION

Not applicable.

CONFLICT OF INTERESTS

The authors declare that there is no conflict of interest regarding the publication of this paper.

GENERATIVE AI STATEMENT

The authors used ChatGPT (OpenAI) solely for English language editing and grammatical correction. Generative AI was not used for content generation, data analysis, or interpretation of the results.

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بندپایان درشت خاکزی و میان خاکزی جنگل‌های حفاظت‌شده هیرکانی: الگوهای فصلی تنوع و ساختار اجتماعی بندپایان در عمق‌های مختلف خاک

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چکیده: جنگل‌های هیرکانی به عنوان یکی از قدیمی‌ترین جنگل‌های جهان در نظر گرفته می‌شوند و به دلیل اقلیم معتدل و مرطوب و خاک‌های حاصلخیز خود، دارای تنوع زیستی غنی هستند. بندپایان ساکن در خاک نقش‌های اساسی در عملکرد اکوسیستم جنگل ایفا می‌کنند و به تجزیه برگ‌ها، چرخه مواد مغذی و حفظ ساختار خاک کمک می‌کنند. این مطالعه تنوع و توزیع عمودی ماکروفون و مزوفون را در جنگل سمسکنده، یک جنگل حفاظت‌شده هیرکانی در شمال ایران، در عمق‌های مختلف خاک (برگ‌های ریخته شده، ۰-۳ سانتی‌متر، ۳-۶ سانتی‌متر) و فصول مختلف در طول یک سال ارزیابی کرد. بیش از ۱۰،۰۰۰ فرد جمع‌آوری شد که نمایانگر گروه‌های تاکسونومیک متنوعی از جمله مورچه‌ها، هزارپاها، دمنف‌ها، کنه‌ها، شبه‌عقرب‌ها و دم‌چنگالان بودند. تراکم جمعیت‌ها و تنوع گونه‌ها به طور مداوم در لایه برگ بیشترین مقدار را داشت و با اوج‌های فصلی واضحی که در طول دوره مطالعه مشاهده شد با افزایش عمق خاک کاهش یافت. مورچه‌ها گروه غالب در ماکروفون بودند، در حالی که کنه‌ها و دمنف‌ها متنوع‌ترین گروه‌های مزوفون بودند. این مطالعه الگوهای فصلی و عمودی واضحی را در تراکم جمعیت و تنوع جوامع جانوری خاک در اکوسیستم جنگل هیرکانی نشان داد. این یافته‌ها اطلاعات پایه‌ای برای حفاظت از تنوع زیستی و نظارت آینده بر جانوران خاک در جنگل‌های هیرکانی فراهم می‌کند. بنابراین، حفظ کیفیت برگ‌ها، حفظ یکپارچگی زیستگاه و نظارت بر تنوع جانوران خاک برای حفظ عملکردهای اکوسیستم جنگل در این منطقه‌ای به عنوان میراث ارزشمند جهانی، بسیار حیاتی است.

واژگان کلیدی: فراوانی، ترکیب جامعه، حفاظت، تنوع جانوری، ایران، خاک‌برگ